

# Improving the monitoring of multi-class pesticides in baby foods using QuEChERS-UHPLC-Q-TOF with automated identification based on MS/MS similarity algorithms

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# 1 Improving the monitoring of multi-class pesticides in baby foods

# 2 using QuEChERS-UHPLC-Q-TOF with automated identification

# **3 based on MS/MS similarity algorithms**

4	Yassine Makni <sup>a</sup> , Thierno Diallo <sup>a, b</sup> ,	Thierry Guérin <sup>c</sup> , Julien Parinet <sup>a,*</sup>
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#### 33 Abstract

A screening method was developed for the multi-residue analysis of pesticides in baby foods 34 using QuEChERS and UHPLC-Q-TOF. For sample preparation, the two-buffered versions of 35 QuEChERS and different purification procedures were studied. False negatives and false 36 positives were determined using different thresholds mentioned in the literature on the 37 retention time and accurate mass measurement detection criteria. To reach unequivocal 38 identification, the fragmentation spectra of the pesticides were used. The information-39 dependant-acquisition (IDA) mode was optimized with a precursor-inclusion list (PIL) to 40 limit the loss of MS/MS data. Then, the experimental fragmentation spectra were compared to 41 those included in a homemade library, by assessing different MS/MS algorithms and 42 similarity scores. The optimised method was validated according to SANTE/11312/2021 43 guidelines. 95% and 73% of the pesticides presented a screening detection limit (SDL) and a 44 limit of identification (LOI)  $\leq 0.1 \text{ mg.kg}^{-1}$ . One plasticizer was found in the investigated 45 samples by a suspect-screening approach. 46

## 47 Keywords

48 LC-HRMS; Screening method; Acquisition mode; Detection; Identification; Method49 validation

#### 50 **1. Introduction**

In 2005, C. Wild introduced the concept of exposome as the total environmental and non-51 genetic exposure sources and factors that an organism undergoes throughout their lifetime, 52 potentially associated with multifactorial disorders (Wild, 2005). In the context of dietary 53 exposure, non-genetic factors to which humans can be exposed include microbiological 54 agents and chemicals, such as pesticides. Pesticides represent a wide-range of bioactive 55 chemical compounds of low molecular weight, intensively used in modern agriculture to 56 control pest organisms and increase crop yields. During the past few decades, their 57 widespread use has caused serious concerns worldwide regarding food quality and safety 58 aspects. Importantly, their application to crops and use in livestock may result in chemical 59 60 pollution in food products (Pang et al., 2021, which exposes the population to a wide array of pesticides occurring in a dynamic range of concentrations. Another important aspect is that 61 62 exposure to mixtures of pesticide residues at different levels may enhance their harmfulness or, on the contrary, reduce it through potentiating effects and synergistic agonistic or 63 64 antagonistic interactions, also known as "cocktail effects" (Lukowicz et al., 2018; Rizzati et al., 2016). As a result, due to their known or potential adverse effects on human health 65 66 (Ntzani et al., 2013), there is a need to assess the risks of chronic exposure to pesticides through diet. 67

The risks to health are even higher in infants below 3 years of age, who represent a more 68 vulnerable group (Nougadère et al., 2020; Aguilera-Luiz et al., 2012). In Europe, the 69 European Food Safety Authority (EFSA) has set maximum residue levels (MRLs) of 70 individual pesticide residues in common animal and plant-based food products for human 71 consumption (European Commission, 2005). Baby foods, a term used to designate mixtures of 72 73 different matrices formulated for infants, including fruits and vegetables, animal and cerealbased food products (Hercegová et al., 2007), follow a standardised MRL of 0.01 mg.kg<sup>-1</sup> 74 (European Commission, 2006). 75

In recent years, multi-residue screening of pesticides has developed significantly with the introduction of new methods, especially regarding sample preparation. The Quick, Easy, Cheap, Efficient, Rugged and Safe (QuEChERS) procedure stands out as one of the most commonly used methods for extraction of multi-class pesticides in food samples (González-Curbelo et al., 2015), including baby foods (Pérez-Ortega et al., 2017; Jia et al., 2014; Aguilera-Luiz et al., 2012). The QuEChERS procedure involves extraction using acetonitrile and partitioning salts, followed by a clean-up step (Anastassiades et al., 2003). Although the
QuEChERS procedure is well suited to extraction of multi-class pesticides with high yields in
various matrices, there is no appropriate one-size-fits-all method. Therefore, prior
optimisation may be required to improve extraction efficiency.

Regarding analytical techniques, gas and reversed-phase liquid chromatography coupled with tandem mass spectrometry (LC/GC-MS/MS) operating in multiple reaction monitoring (MRM) mode are the most developed technique for routine pesticide determination (Villaverde et al., 2016; Alder et al., 2006). Although these methods are very sensitive and robust, a major limitation is that they are limited to a predetermined list of targeted compounds to be searched *a priori*. These targeted approaches are blind to any compound outside this list.

The rise of high-resolution mass spectrometry (HRMS) has made it possible to overcome this 93 limitation. Unlimited determination of compounds without a pre-defined list has become 94 possible using high-resolution mass spectrometers such as the time-of-flight (TOF) or 95 Orbitrap instruments operating in full-scan MS mode (Weng et al., 2020). This feature 96 97 improved the concept of screening methods and allowed for the development of suspect and non-targeted screening approaches, which extend the analytical scope (Dubocq et al., 2021; 98 99 Phillips et al., 2018). Furthermore, retrospective analysis of acquired full-MS data is possible 100 when a new contaminant was not characterised at the time of injection (Gómez-Pérez et al., 101 2015).

102 Detection of pesticides in food using HRMS operating in full-scan mode is based on the retention time (t<sub>r</sub>) and accurate mass measurement of a precursor ion. To date, there is no 103 established consensus on the thresholds to be used for these criteria, even though analytical 104 guidelines (SANTE, 2021; FDA, 2015) provide various recommendations. For instance, 105 SANTE guidelines recommend a t<sub>r</sub> tolerance of  $\pm 0.1$  min and accurate mass measurement  $\pm$ 106 107 5 ppm for screening of pesticides in food, whereas some authors prefer to use wider thresholds (Wang et al., 2019; Pérez-Ortega et al., 2017; Luz Gómez-Pérez et al., 2015). This 108 109 is a crucial parameter to consider as the data processing parameters may influence the balance 110 between false negatives and false positives (López et al., 2014). Given this consideration, an assessment on the thresholds to be used should be carried out at the time of method 111 development, to provide for efficient and reliable automated-based data processing, which is a 112 key for implementation of HRMS in routine use. 113

To reach a higher degree of confidence and limit the occurrence of false positives, screening 114 of pesticides may also include MS/MS data for unequivocal identification (Schymanski et al., 115 2014). In HRMS, full-scan MS mode is often followed by MS/MS data acquisition in the 116 same run-analysis, which can be performed in two ways: full-scan mode followed by either 117 information-dependent-acquisition (IDA) or data-independent acquisition (DIA). IDA mode 118 provides MS/MS spectra of high quality, but all the analytes of interest are not systematically 119 fragmented (Guo et al., 2020). To overcome this limitation, a precursor-inclusion list (PIL) 120 can be included to prioritize the compounds to be fragmented in targeted analysis. Another 121 122 option is the use of DIA mode, in which all the precursor ions are fragmented, providing more complex spectra (Diallo et al., 2022. There are no requirements on the mode to be used for 123 124 identification. In guidelines (SANTE, 2021; FDA, 2015), identification entails the use of t<sub>r</sub> and accurate mass measurement of a precursor ion and corresponding product ion(s). 125 126 However, not all instrumental software supplies a mass error (ppm) on product ions, and the identification then relies on the similarity of the whole MS/MS spectrum of a compound, 127 128 against its reference spectrum included in a home-made library. The degree of similarity is displayed by scores (%) measured using MS/MS search algorithms. Nevertheless, this 129 130 approach is not covered by the guidelines, which may be detrimental if this is the only option provided by the software. In the literature, few studies have reported the use of automated 131 identification based on MS/MS spectra similarity for multi-residue screening of pesticides in 132 food (Wang et al., 2016). 133

The aim of this work was to develop a simple and reliable method for the screening of 204 134 pesticides in baby food samples using UHPLC-Q-TOF analysis. The sample preparation 135 conditions were optimised based on a QuEChERS protocol that can be used for baby food of 136 137 heterogeneous and varied composition. To the best of our knowledge, this study is the first evaluating different QuEChERS methods and clean-up procedures to analyse multi-class 138 139 pesticides in such heterogeneous matrices as baby foods. False negatives and false positives were determined using different thresholds mentioned in the literature on the t<sub>r</sub> and accurate 140 141 mass measurement criteria.. Another part of the work focused on the identification of the pesticides. Thus, the impact of a PIL during data acquisition was evaluated on the 142 143 performances of identification. Furthermore, this study is the first to evaluate the impact of different MS/MS search algorithms and similarity scores, regarding the performances of 144 145 identification of multi-class pesticides in baby food samples. The method was validated based 146 on the SANTE guidelines (SANTE, 2021). The screening detection limit (SDL) of each of the studied pesticides was thus established. Although there is no requirement for meeting identification criteria, the limit of identification (LOI) was also considered as a validation parameter. The versatility of the method was tested using an expanded list of organic contaminants in suspect-screening analysis.

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### **2. Material and Methods**

#### 153 2.1. Chemicals and reagents

A kit of 204 LC pesticide standards, divided into ten mixtures (100 mg.L<sup>-1</sup> of each compound 154 in either acetonitrile (MeCN) or methanol (MeOH)), was obtained from Restek Corporation 155 (Ulis, France). HPLC-MS-grade MeCN and formic acid (FA) were obtained from 156 Thermofisher Scientific Corporation (Illkirch, France). Ultrapure water (18.2 MQ.cm) used 157 for reagent and sample preparation was obtained through a Millipak 40 Gamma Gold Filter 158 (pore size 0.22 µm) and using a Milli-Q Ultrapure and Pure water system purchased from 159 Merck Millipore (Saint-Quentin-en-Yvelines, France). A QuEChERS extraction kit, based on 160 the AOAC Method and containing 4.0 g of magnesium sulphate (MgSO<sub>4</sub>) and 1.0 g of sodium 161 acetate, was purchased from Agilent (Santa Clara, CA, USA). A QuEChERS extraction kit 162 based on EN Method 15662 and containing 4.0 g of MgSO<sub>4</sub>, 1.0 g of sodium chloride (NaCl), 163 1.0 g of sodium citrate and 0.5 g of sodium citrate sesquihydrate, was purchased from Agilent 164 (Santa Clara, CA, USA). A Dispersive solid phase extraction (d-SPE) kit containing 150 mg 165 of primary and secondary amines (PSA) and 900 mg of MgSO<sub>4</sub> was purchased from Agilent 166 (Santa Clara, CA, USA). A kit of d-SPE containing 400 mg of PSA, 1200 mg of MgSO<sub>4</sub>, 400 167 mg of octadecylsilyl silica (C18) and 45 mg of Graphitized Carbon Black (GCB) was 168 obtained from Agilent (Santa Clara, CA, USA). An SPE Oasis Prime HLB purification 169 cartridge was acquired from Waters (Milford, MA, USA). 170

171

#### 172 2.2. Equipment and apparatus

173 IKA Utra-Turrax T25 mixer used for sample extraction was obtained from Retsch
174 Technology GmbH (Haa, Germany). A Genie 2 vortex used for sample extraction procedure
175 was purchased from Scientific Industries (Bohemia, USA), and an Eppendorf 5810 centrifuge
176 was purchased from Eppendorf (Hamburg, Germany).

The UHPLC-Q-TOF system used for analysis consisted of a Dionex UltiMate-3000 system
with a binary pump and autosampler (ThermoFisher Scientific, Waltham, USA), coupled with
a Q-TOF mass spectrometer (TripleTOF 5600<sup>TM</sup>, AB Sciex, Darmstadt, Germany) equipped
with an electrospray ionization source (ESI).

181

#### 182 2.3. Preparation of standard solutions

The ten different mixtures (#1 to #10), 100 mg.L<sup>-1</sup>, were transferred and stored in screwcapped amber glass tubes at  $-20^{\circ}$ C. There were 13 pesticides in mix #1, 16 pesticides in #2, 38 in #3, 63 in #4, 30 in #5, 28 in #6, 7 in #7, 1 in #8, 7 in #9, and finally 1 pesticide in mix #10. A mixture of the 204 pesticides (2 mg.L<sup>-1</sup>) was prepared by a suitable dilution of the 10 individual mixtures with MeCN, and stored in screw-capped amber glass tubes at  $-20^{\circ}$ C. The composition of each mix is supplied in electronic supplementary materials (ESM) (Table S1).

#### 189 2.4. Pesticides selected

The 204 pesticides included in this study were selected in view of their worldwide occurrence 190 in food samples. The wide range of physicochemical properties of these pesticides make them 191 192 a relevant model for a multi-residue method. Among these compounds, only the 63 pesticides belonging to mix #4 (Table S1) were selected for the fortification of baby food in the sample 193 194 preparation optimisation-stage. They were selected as they cover the dynamic range of polarity (LogP) of the rest of the studied pesticides. The complete mixture of 204 pesticides 195 196 was used for the fortification of the baby food samples for the following parts: (i) impact of different criteria thresholds on the performances of the method, (ii) impact of a PIL and 197 198 different MS/MS algorithms on the performances of the method, and (iii) method validation.

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#### 200 2.5. Matrices selected

Fourteen baby food commodities of various matrix compositions were purchased from local 201 markets (Maisons-Alfort, France), and stored in a freezer at -20 °C. The composition of each 202 baby food is supplied in ESM (Table S2). Four of the fourteen baby food samples (#1 to #4), 203 representative of the remaining foods in terms of matrix composition, were selected for 204 optimisation of sample preparation. The ten remaining baby food samples (#5 to #14) were 205 used for the other parts of the study: (i) impact of different criteria thresholds on the 206 performances of detection, (ii) impact of a PIL and different MS/MS algorithms on the 207 performances of identification, (iii) method validation, and (iv) suspect screening analysis. 208

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#### 210 2.6. UHPLC-Q-TOF analysis

Reversed-phase gradient liquid chromatography separation was performed using an 211 Aqua<sup>®</sup> C18 column (150 x 2 mm, 3-µm particle size, 125 Å pore diameter; Phenomenex Inc, 212 USA) equipped with an Aqua<sup>®</sup> pre-column (4.0 mm x 2.0 mm, 3-µm particle size, 213 Phenomenex Inc, USA). The column temperature was thermostatted at 50°C. The mobile 214 phase composition consisted of ultrapure water (eluent A) and MeCN (eluent B), both 215 216 containing 0.01% of FA. The gradient profile was started with 10% B (0 min); increased linearly to 100% B for 9 min (0-9 min); fixed for 3 min (9-12 min), returned to the initial 217 218 conditions for 0.1 min (12.1 min), and then followed by a re-equilibration step for 3.9 min. 219 The total run time was 16 min. The autosampler was thermo-regulated at 15°C. The injection 220 volume was 10  $\mu$ L and the gradient flow rate was constant and set at 0.5 mL.min<sup>-1</sup>.

221

Before each batch, the TripleTOF 5600<sup>TM</sup> system was systematically tuned and calibrated 222 manually using the APCI positive calibration solution (AB Sciex technology) in TOF MS and 223 TOF MS/MS in high-resolution mode. Furthermore, the automated calibration device system 224 225 (CDS) was scheduled to perform automatic calibration every five samples during the runanalysis. ESI+ parameters were as follows: ion spray voltage floating (ISVF) was set at 5.5 226 kV, source temperature (TEM) 500°C, curtain gas (CUR) 30 psi, nebulising gas (GS1) 40 psi, 227 and heater gas (GS2) 45 psi. In this study, two methods based on IDA mode were carried out 228 using the UHPLC-Q-TOF system: 229

- Method A was a combination of full scan TOF-MS / IDA mode. Five experiments were 230 conducted in a single-run analysis. The first experiment was a full-scan TOF-MS acquisition 231 that covered a mass range of 50-1000 Da. The declustering potential was set at 80 eV and the 232 collision energy (CE) was set at 10 eV. The accumulation time was set at 490 ms. The 233 experiments 2-5 consisted in MS/MS data acquisition using IDA mode. The accumulation 234 235 time was set at 100 ms for each IDA experiment, and the CE was set at 40 eV with a spread of  $\pm$  20 eV (20, 40 and 60 eV) in high-resolution mode. The collision-induced dissociation gas 236 used was nitrogen. Dynamic background subtraction was turned on. The total cycle time was 237 about 900 ms. Considering that the chromatographic peak width was approximatively 15 s, 238 the obtained number of points per peak was about 16-17 which is quite enough according to 239 the SANTE guidelines. 240
- 241

- Method B consisted in the same full scan TOF-MS / IDA acquisition mode than Method A, with the same parameters. However, in this method, a PIL of the 204-targeted pesticides was implemented to avoind the possible non-fragmentation of analytes. PIL comprised the 204 precursor ion masses in forms of [M+H]<sup>+</sup>, [M+Na]<sup>+</sup>, [M+NH4]<sup>+</sup>, or insource fragments, as well as their corresponding chromatographic t<sub>r</sub>. These informations are supplied in ESM (Table S1).

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Data acquisition was performed using Analyst<sup>®</sup> software (version 1.7) of AB Sciex (Les Ulis,
France). The resulting extracted ion chromatograms (XICs) were reconstructed using
PeakView<sup>®</sup> software, version 2.2 comprising the MasterView<sup>TM</sup> module of AB Sciex.

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#### 253 2.7. Development of a homemade MS/MS library

A homemade MS/MS library of 204 pesticides was built using the convenient LibraryView<sup>TM</sup> software, version 1.0.3 of AB Sciex. The library was built by injecting separately each of the ten mixtures diluted to  $0.2 \text{ mg.L}^{-1}$  in MeCN. Exact mass, the CAS number and the MS/MS fragmentation spectrum of each of the studied pesticides were recorded. The fragmentation spectra included in the homemade library are considered as reference spectra.

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#### 261 2.8. *Optimisation of sample preparation*

Optimisation of extraction was performed comparing the two-buffered variants of QuEChERS, the AOAC 2007.01 (Lehotay, 2007) and CEN 15662 (Anastassiades et al., 2007) methods. In this study, these methods were modified including a hydrating strategy and without extraction solvent acidification.

Sample preparation optimisation is presented in Fig. 1. Before use, the samples were thawed 266 at 4°C overnight. No sample pre-preparation was required. For the enrichment experiment, 15 267  $\mu$ L of mix #4 diluted to 2 mg.L<sup>-1</sup> was added to 50 mL-polypropylene centrifuge tubes, which 268 contained 3.0 g of blank sample  $\pm$  0.2 g, to provide a spiking level at 0.01 mg.kg<sup>-1</sup> of 269 pesticides. A volume of 7 mL of ultrapure water was added to each tube, and then ground 270 271 using an IKA Ultra-Turrax mixer for 3 min at 5000 rpm. Then, 10 mL of MeCN was added to each centrifuge tube to perform the extraction of pesticides, which were vortexed again for 1 272 273 min. Then, either AOAC or CEN extraction salts were added to the fortified samples, and the resulting mixture was centrifuged for 5 min at 4000 rpm. 1 mL of the resulted supernatant was 274

275	deposited into a 2 mL-amber vial for injection in the UHPLC-Q-TOF system operating with
276	Method A.
277	Sample preparation optimisation also included a comparison of different protocols of
278	purification after the extraction-stage, in addition to the protocol without purification (Fig. 1).
279	When a purification step was performed, 6 mL of the supernatant was recovered and
280	deposited/percolated into either:
281	- A 15-mL centrifuge tube containing 900 mg of MgSO <sub>4</sub> and 150 mg of PSA;
282	- A 15-mL centrifuge tube containing 1200 mg of MgSO <sub>4</sub> , 400 mg of PSA, 400 mg of
283	C18 and 45 mg of GCB;
284	- An Oasis HLB cartridge, without prior conditioning.
285	
286	Finally, 1 mL of the supernatant was deposited into a 2 mL-amber vial prior to injection in the
287	UHPLC-Q-TOF system operating with Method A.
288	All protocols (Fig. 1) were performed in triplicate. The assessment was carried out
289	considering the false negative rates of each procedure. A pesticide was considered a false
290	negative when a pesticide did not fit a $t_r$ variation $\pm 0.1$ min and/or accurate mass
291	measurement± 5 pm (SANTE, 2021).
292	

293 2.9. Impact of different detection criteria thresholds on the performances of
294 detection

In this section, the term "detection" was used to define the screening of pesticides using  $t_r$  and accurate mass measurement criteria on the precursor ions. In this study, an assessment of the criteria of detection was carried out using the following thresholds: (i)  $t_r$  variation  $\pm$  0.1 min, accurate mass measurement  $\pm$  5 ppm (SANTE, 2021); (ii)  $t_r$  variation  $\pm$  0.5 min, accurate mass measurement  $\pm$  5 ppm (Wang et al., 2019; FDA, 2015); (iii)  $t_r$  variation  $\pm$  0.25 min, accurate mass measurement  $\pm$  10 ppm (Pérez-Ortega et al., 2017).

Each of the studied thresholds was evaluated regarding the false negative and false positive rates obtained. A false negative was defined when a pesticide did not fit the expected thresholds for  $t_r$  and/or mass measurement criteria. In contrast, a false positive was defined when the software reported a positive hit, i.e., when a pesticide fitted both  $t_r$  and accurate mass measurement criteria thresholds, in the non-spiked samples. The same list of 204 pesticides was used for the calculation of false negatives and false positives.

# 307 2.10. Impact of a precursor ion list and different MS/MS similarity algorithms on the 308 performances of identification.

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310 In this section, the term "identification" was used to define the screening of the 204 pesticides using  $t_r$  variation  $\pm 0.1$  min and accurate mass measurement  $\pm 5$  ppm on the precursor ion, as 311 well as the characteristic MS/MS fragmentation spectra. In this part of the work, the baby 312 food samples were acquired using Method A and Method B, in order to evaluate the impact of 313 the PIL on the performances. Then, the generated files were processed using MasterView<sup>TM</sup> 314 315 software. The experimental fragmentation spectra obtained were compared to those included in the homemade library. This comparison was automatically computed by the software using 316 317 MS/MS search algorithms. Three algorithms are available in the software, and were assessed in this study: 318

- The *purity algorithm* (PUR) measures the similarity between the experimental and
   reference spectra. All fragment ions from both spectra and their relative abundance are
   used for calculation;
- The *fit algorithm* (FIT) measures the degree to which the library spectrum is contained
   in the experimental spectrum, disregarding additional fragment ions present in the
   experimental spectrum but absent in the reference spectrum. All the expected fragment
   ions and their relative abundance are used for calculation;
- The *Reverse Fit* algorithm (RFIT) measures the degree to which the experimental
   spectrum is contained in the library spectrum, disregarding the fragment ions present
   in the library spectrum but absent in the experimental spectrum. All the experimental
   fragment ions and their relative abundance are used for calculation.

These algorithms automatically calculate a degree of confidence characterised by a similarity score (%), allowing for fast and reliable identification, as illustrated in ESM (Fig. S1). Two similarity scores: (i) >50% and (ii) >70%, used elsewhere (Ogawa et al., 2020; Wang et al., 2016), were considered as the thresholds to be fulfilled. A pesticide that was not fitting the above mentionned criteria or not fragmented was reported a false negative.

335

336 2.11. Method validation

The method was validated according to the SANTE guidelines for screening methods. Method 337 validation was conducted at 0.01 and 0.1 mg.kg<sup>-1</sup> to establish the screening detection limits 338 (SDL) (SANTE, 2021) and the limits of identification (LOI). Experiments were carried out in 339 triplicate for repeatability assessment. Each replicate was injected in a single batch, on three 340 different days, to include inter-batch and instrumental over-time variabilities.. The SDL was 341 established as the lowest concentration for which a compound was detected following (i)  $t_r$ 342 variation  $\pm$  0.1 min and (ii) accurate mass measurement  $\pm$  5 ppm, in at least 95% of the 343 samples (SANTE, 2021), i.e., in at least 29 out of the 30 baby food samples. In addition, the 344 345 LOI was established as the lowest concentration for which a pesticide was correctly identified in at least 95% of the samples, i.e., in at least 29 out of the 30 baby food samples. The LOI 346 used the same criteria as for the SDL, plus a comparison between experimental and reference 347 fragmentation spectra using FIT algorithm with a similarity score >70%. 348

349

## 350 **3. Results and discussion**

#### 351 *3.1. Sample preparation optimisation*

The sample preparation was studied and optimised regarding different extraction and purification procedures as explained in Section 2.7. The results for the four baby foods are displayed in Fig. 2.

355

356 When using AOAC mixture salts, false negative rates ranged from 22–32%, 19–33%, 21–32% and 13-22% in baby food 1 to 4, respectively regardless of the purification procedure used. 357 Then, when using CEN mixture salts, false negative rates varied from 29-38%, 32-39%, 29-358 38% and 21–29% in baby food 1 to 4, respectively regardless of the purification procedure used. 359 According to the results, the extraction of pesticides appeared to be more effective using AOAC 360 salts. Regarding the purification following AOAC-based extraction, false negatives varied from 361 19-27%. 13-22%, 22-33% purification. 362 and 19-32% using no MgSO<sub>4</sub>/PSA, MgSO<sub>4</sub>/PSA/C18/GCB and the Oasis PRIME HLB cartridge, respectively. Regarding the 363 purification following CEN-based extraction, false negatives varied from 23-34%, 21-32%, 21-364 36% and 29-39% using no purification, MgSO4/PSA, MgSO4/PSA/C18/GCB and the Oasis 365 PRIME HLB cartridge, respectively. According to the results, purification using MgSO<sub>4</sub>/PSA 366 was the most effective to clear the extracts achieved with the AOAC or CEN methods. 367

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The results obtained separately from the four baby foods (Fig. 2) were grouped together to 369 increase the statistical power and to evaluate the overall performances, irrespective of the matrix 370 composition. The results are displayed in ESM (Fig. S2). Average false negative rates ranged 371 from 19-30% for the AOAC method against 28-36% for the CEN method, regardless of the 372 purification procedure. Performances of detection were better when using the AOAC method, 373 which was statistically confirmed (two-way ANOVA: p-value < 0.01). Therefore, the acetate-374 buffered version of QuEChERS was retained for the extraction of pesticides. This was consistent 375 with other studies in which AOAC mixture salts were used for screening of multi-class 376 377 contaminants in baby foods (Pérez-Ortega et al., 2017; Aguilera-Luiz et al., 2012).

378

379 Then, the four purification protocols were assessed for the retained AOAC method. On average, false negative rates were 23%, 19%, 30% and 27% when using the procedure without 380 381 purification, MgSO<sub>4</sub>/PSA, MgSO<sub>4</sub>/PSA/C18/GCB and the Oasis PRIME HLB cartridge, respectively. Standard deviations (SD) were equivalent overall. Purification using MgSO<sub>4</sub>/PSA 382 383 appeared to be the most efficient protocol to remove matrix interferents and improve compound detectability. MgSO<sub>4</sub> acts as a desiccant as it dries out the water to improve the partitioning 384 385 process (Li et al., 2021), whereas PSA is commonly used in various food as a sorbent to remove 386 co-extractives sugars, organic acids, fatty acids and certain pigments (Walorczyk et al., 2015). Nevertheless, there was no significant difference between this protocol and the protocol without 387 a purification step. Thus, to be able to select the most suitable protocol (MgSO<sub>4</sub>/PSA vs no 388 purification), the signal-to-noise ratio (S/N ratio) was considered. The results are supplied in 389 ESM (Fig. S3). The average S/N ratios were 371 and 476 without purification and with 390 MgSO<sub>4</sub>/PSA, respectively. On average, pesticide signals were better when using a clean-up step 391 with MgSO<sub>4</sub>/PSA, which was statistically confirmed (t-test, *p*-value  $\leq 0.05$ ). Purification using 392 MgSO<sub>4</sub>/PSA was the most efficient protocol to improve the S/N of the pesticides and was 393 therefore retained for the method. The lower S/N ratio in the samples without purification can be 394 explained by the higher amount of co-extracted compounds (sugars, organic acids, fatty acids, 395 396 pigments, etc.), which could lead to signal suppression of pesticides during the ionisation process (Petrarca et al., 2022). The use of MgSO<sub>4</sub>/PSA as a purification method was also conducted by 397 Pérez-Ortega et al., (2017) on a list of more than 600 contaminants in various matrices, including 398 baby foods. 399

400

401 This sample preparation protocol could be further transposed to matrices other than baby foods,402 only requiring minor adjustments, as baby foods are mixtures of various food commodities of

403 plant and animal origin. Moreover, the AOAC 2007.01 QuEChERS method as well as 404 MgSO<sub>4</sub>/PSA are widely used for multi-residue screening of organic pollutants in various food 405 samples (Kim et al., 2019; Musarurwa et al., 2019). Furthermore, the low-starting sample 406 amount (3 g) coupled to a hydrating strategy enable its further application to matrices with low-407 water content ( $\leq 25\%$ ), for which hydration is crucial to ensure effective extraction (Zhang et al., 408 2019).

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# 410 3.2. Impact of different detection criteria thresholds on the performances of the 411 method

There is no established consensus on the thresholds to be used for t<sub>r</sub> variation and accurate 412 mass measurement criteria. Moreover, an assessment of these detection criteria is required to 413 obtain the best compromise between the false positives and false negatives reported by the 414 415 software, as explained by (Gómez-Ramos et al., 2013). As a result, different thresholds detailed in section 2.9. were evaluated. Results regarding the false negative rates are displayed 416 in Table 1. When  $t_r \pm 0.1$  min and accuracy mass measurement  $\pm 5$  ppm were used, the false 417 negative rates ranged from 13% to 21% with an average of 16% at 0.01 mg.kg<sup>-1</sup>, and from 3% 418 to 9% with an average of 5% at 0.1 mg.kg<sup>-1</sup>. When varying only the t<sub>r</sub> threshold to  $\pm$  0.5 min, 419 the false negative rates ranged from 13% to 20% with an average of 16% at 0.01 mg.kg<sup>-1</sup>, and 420 from 3% to 9% with an average of 5% at 0.1 mg.kg<sup>-1</sup>. Finally, when setting the  $t_r$  threshold to 421  $\pm$  0.25 min and accuracy mass measurement to  $\pm$  10 ppm, the false negative rates varied from 422 9% to 13% with an average of 11% at 0.01 mg.kg<sup>-1</sup>, and from 3% to 4% with an average of 423 3% at 0.1 mg.kg<sup>-1</sup>. These results demonstrate that a wider t<sub>r</sub> variation with unchanged accurate 424 mass measurement (5 ppm) did not impact the average detectability, regardless of the 425 426 concentration level. However, a larger accurate mass measurement threshold (± 10 ppm) slightly improved pesticide detection, especially at 0.01 mg.kg<sup>-1</sup>. These findings are consistent 427 428 with the study of López et al. (2014) which shown the same trends on 199 pesticides in fruits and vegetables. Otherwise, SDs for each baby food were all  $\leq 7\%$ , suggesting relatively good 429 inter-batch repeatability. The total inter-batch and inter-matrix SDs were  $\leq 4\%$ , regardless of 430 the thresholds used, suggesting good comprehensiveness of the method. 431

432

False positives should also be considered as a criterion for evaluating the effectiveness of a
screening method in HRMS. As there is no requirements in official guidelines (SANTE, 2021;

FDA, 2015) on the false positive rates not to be exceeded, the proposed automated screening 435 method must be able to provide false positive rates as low as possible, to avoid time-436 consuming manual inspections. Results on the false positives are displayed in ESM (Table 437 S3). When using  $t_r \pm 0.1$  min and accuracy mass measurement of  $\pm 5$  ppm, the false positive 438 rates ranged from 3% to 6% with an average of 5%, and an SD of 2%. When increasing the tr 439 threshold to  $\pm 0.5$  min, while keeping the same accurate mass measurement of  $\pm 5$  ppm, the 440 false positive rates varied from 4% to 8% with an average of 6%, and an SD of 2%. Finally, 441 when  $t_r \pm 0.25$  min and accuracy mass measurement  $\pm 10$  ppm were used, the false positive 442 rates varied from 9% to 15%, with an average of 13%, and an SD of 5%. Thus, the false 443 positive rates depended more heavily on the accurate mass measurement used than on the t<sub>r</sub> 444 445 variation threshold. These results are consistent with the findings of Pérez-Ortega et al. (2017) and López et al. (2014) regarding the influence of wider accurate mass measurement on false 446 447 positives. However, the results were different considering the tr criteria, as these authors showed higher false positives with the widening of the t<sub>r</sub> threshold (Pérez-Ortega et al., 2017; 448 449 López et al., 2014). In conclusion, the thresholds recommended in SANTE guidelines ( $t_r \pm 0.1$ min, accurate mass measurement  $\pm$  5 ppm) (SANTE, 2021) allowed us to obtain the best 450 451 balance between false negative and false positive rates, and were therefore selected for 452 pesticide detection.

453

To the best of our knowledge, this study is the first to provide both false negative and false 454 positive rate results for multi-residue screening of pesticides in baby food samples. However, 455 data of this type are available for similar related matrices, such as fruits and vegetables. For 456 instance, López et al. (2014) proposed a method for the screening of almost 200 pesticides in 457 a set of 22 fruit and vegetable samples using a QuEChERS procedure followed by UHPLC-Q-458 TOF (ESI+) analysis. They were able to detect from 61% to 94% of pesticides with an 459 average of 84% at 0.01 mg.kg<sup>-1</sup>, i.e., 16% false negatives, when using  $t_r$  of  $\pm$  0.2 min and 460 accurate mass measurement of  $\pm$  5 ppm. The performances of detection in the present study 461 462 and in the one of López et al. (2014) were equivalent (16% false negatives), but considering a tighter t<sub>r</sub> threshold. On the other hand, they found up to 15 false positives on average against a 463 library of 504 entries, i.e., a false positive rate of about 3%, which was equivalent to that in 464 the present work (5%). In the study conducted by López et al. (2014), a QuEChERS 465 extraction-based protocol was performed using equivalent sample amounts and extraction 466 solvent volumes, making it possible to obtain 1 g of sample per mL of final extract. In 467 468 contrast with the present study, the starting sample mass (3 g) was lower than the

469 subsequently added MeCN (10 mL), providing about 0.3 g of sample per mL of final extract, 470 which may decrease co-extractive compounds but also increase analyte detection limits. The 471 overall method provided satisfactory detection performances with low-related variability and 472 low false positive rate, even at low-level of concentration and given the complex nature of 473 baby food samples.

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# 475 3.3. Impact of a PIL and different MS/MS algorithms on the performances of the 476 method

To achieve a simple and reliable automated screening method with unequivocal identification (level 1 confidence), it is necessary to consider both information regarding precursor ion and its characteristic fragment ions (Schymanski et al., 2014). For this purpose, the methods detailed in section 2.5 and the protocols detailed in section 2.10 were evaluated regarding pesticide identification. The results are displayed in Table 2.

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Method A provided false negative rates from 50% to 55% at 0.01 mg.kg<sup>-1</sup>, and from 26% to 483 28% at 0.1 mg.kg<sup>-1</sup>, depending on the algorithm-score combination used. The false negatives 484 485 were higher than those found at detection-stage (Table 1), which is consistent because for identification, additional parameter, comparison between experimental and library 486 fragmentation spectra, is taken into account. Method B provided false negative rates from 487 29% to 42% at 0.01 mg.kg<sup>-1</sup>, and from 13% to 17% at 0.1 mg.kg<sup>-1</sup>, depending on the 488 algorithm-score combination used. Method B always provided the lowest false negative rates 489 at both concentration-levels, which can be explained by the use of a PIL, giving priority to the 490 pesticides to be fragmented and limiting the loss of MS/MS data. Regarding the fragmentation 491 triggering rate, it was found that 31% and 19% of the detected pesticides were not fragmented 492 at 0.01 and 0.1 mg.kg<sup>-1</sup>, respectively with **Method A**. When using **Method B**, the rates of 493 detected but unfragmented pesticides dropped on average to 8% and 7% at 0.01 mg.kg<sup>-1</sup> and 494 0.1 mg.kg<sup>-1</sup>, respectively. Interestingly, the PIL did not completely avoid the loss of 495 information on MS/MS data, preventing these compounds from being unambiguously 496 identified. However, it considerably overcame this limitation when compared with Method 497 A. As a result, Method B was retained for data acquisition. 498

Regarding algorithm-score combinations (**Method B**), the false negative rates ranged from 36% at 0.01 mg.kg<sup>-1</sup> to 14% at 0.1 mg.kg<sup>-1</sup> using PUR, from 33% at 0.01 mg.kg<sup>-1</sup> to 14% at 0.1 mg.kg<sup>-1</sup> using FIT, and from 29% at 0.01 mg.kg<sup>-1</sup> to 13% at 0.1 mg.kg<sup>-1</sup> using RFIT, at a

similarity score >50%. With a more restrictive similarity score (>70%), the false negative 502 rates varied from 42% at 0.01 mg.kg<sup>-1</sup> to 17% at 0.1 mg.kg<sup>-1</sup> using PUR, from 37% at 0.01 503 mg.kg<sup>-1</sup> to 15% at 0.1 mg.kg<sup>-1</sup> using FIT, and from 32% at 0.01 mg.kg<sup>-1</sup> to 14% at 0.1 mg.kg<sup>-1</sup> 504 using RFIT. This demonstrated that pesticide identification was strongly impacted by the 505 concentration-level of analytes in the baby food samples. This was expected, since a higher 506 concentration-level of analyte in the sample leads to better response of precursor and fragment 507 ions, enhancing pesticide identification. Regarding the different algorithms, the results 508 showed equivalent performances at 0.1 mg.kg<sup>-1</sup> regardless of the similarity score used. With 509 lower concentrations, the PUR and FIT algorithms provided equivalent results with a 510 511 similarity score >50%, whereas RFIT made it possible to slightly improve the performances. When using a similarity score >70%, the PUR algorithm showed the worst performances and 512 RFIT the best. Nevertheless, RFIT is the least restrictive algorithm, because even if expected 513 514 fragment ions are not present in the experimental spectrum, they are not taken into account in the calculation.. On the other hand, PUR is the most restrictive algorithm, as it uses the 515 516 fragment ions of both unknown and reference spectra, and their relative abundance for similarity calculation. This explains why this algorithm provided the highest false negative 517 518 rates at low-concentration. The FIT algorithm appeared to be a compromise as it provided results close to those obtained using RFIT, while providing a higher level of confidence. The 519 FIT algorithm presents the advantage of considering the entire expected fragmentation 520 spectrum of the library in the calculation, excluding the interfering fragment ions that can be 521 found in the experimental spectrum. As an example, the pesticide epoxiconazole ([M+H]<sup>+</sup>, 522 m/z 330.08039, t<sub>R</sub> = 6.36 min) was identified with a similarity score of 100% using the FIT 523 algorithm, but with lower scores (67%) using the PUR and RFIT algorithms due to 524 unexpected fragment ions (Fig. S4). Thus, the RFIT and PUR algorithms were both discarded 525 in favour of the FIT algorithm for the identification of pesticides. When using Method B and 526 the FIT algorithm, false negative rates were close either using similarity score >50% or >70%527 (Table 2). Thus, a similarity score >70% was retained for the method, as it provides greater 528 529 confidence in the automated identification process.

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The optimised parameters of identification enabled us to discard all the false positives that were initially detected with only  $t_r$  variation  $\pm 0.1$  min and accurate mass measurement  $\pm 5$ ppm. However, the fungicide difenoconazole ([M+H]<sup>+</sup>, *m/z* 406.07197,  $t_r = 7.31$  min) initially

considered a false positive, was confirmed to be a true positive at the identification-stage. It

was found in non-spiked baby foods 5 and 6. Fig. 3 shows the unexpected identification of difenoconazole in the non-spiked baby food 5. The mass error was 0.2 ppm, the  $t_r$  variation was 0 min, the chlorinated-type isotope ratio difference was about 3%, and the similarity score of identification was about 99% using the FIT algorithm. This demonstrates the reliability of the developed automated identification method, and how important MS/MS fragmentation spectra are for unequivocal identification (Schymanski et al., 2014).

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Considering the literature, López et al. (2014) used a QuEChERS-UHPLC-Q-TOF method to 543 544 screen almost 200 pesticides in fruits and vegetables. As an identification strategy, the authors used  $t_r$  variation  $\pm 0.1$  min and accurate mass measurement  $\pm 5$  ppm on the precursor ion, plus 545 accurate mass measurement  $\pm$  10 ppm on corresponding product ion(s). The average 546 identification rate was 64% at 0.01 mg.kg<sup>-1</sup>, which was equivalent to our results (37% false 547 negative rate, i.e., 63% identification rate). Despite the demanding nature of the identification 548 approach proposed in this study and the complexity of the baby food matrices, the method 549 550 showed satisfactory performances for a large-scale screening purpose, even at low-level of concentration 551

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#### 554 *3.4. Method validation*

The last part of this work consisted in validation of the complete optimised method, inspired 555 by SANTE guidelines for screening methods (SANTE, 2021). SDL and LOI, detailed in 556 section 2.11., were used as criteria of validation. However, as difenoconazole was found in 557 baby foods 5 and 6, these samples were discarded. The results are displayed in ESM (Table 558 S4) and are summarized in ESM (Fig. S5). Regarding the SDL, it was established at 0.01 559 mg.kg<sup>-1</sup> for 141 pesticides (71%), and at 0.1 mg.kg<sup>-1</sup> for 48 pesticides (24%). Overall, 95% of 560 the pesticides had an SDL  $\leq 0.1 \text{ mg.kg}^{-1}$ . These validation results demonstrated the very good 561 detection capabilities of the method, and its ability to act as an alert system if a pesticide 562 found in a baby food sample significantly exceeds the MRL. However, 10 pesticides (5%) 563 were not detected in at least 29 of the 30 baby food samples. Therefore, their SDL was 564 established as >  $0.1 \text{ mg.kg}^{-1}$ . 565

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Regarding the LOI, it was established at 0.01 mg.kg<sup>-1</sup> for 77 pesticides (39%), and at 0.1 mg.kg<sup>-1</sup> for 68 pesticides (34%). Overall, 73% of the pesticides had an LOI  $\leq$  0.1 mg.kg<sup>-1</sup>.

However, 54 pesticides (27%) could not be properly identified in at least 29 of the 30 baby food samples. Therefore, their LOI was established as > 0.1 mg.kg<sup>-1</sup> (Fig. S5).

In the literature, Luz Gómez-Pérez et al. (2015) used  $t_r$  variation  $\pm 0.5$  min and mass error < 5571 ppm on the precursor and one of its fragment ions, for the screening of more than 300 572 pesticides in baby foods. The LOI was established at 0.01 mg.kg<sup>-1</sup> for about 61% of the 573 compounds. In this study, the authors explained that analytes at low-concentration in complex 574 matrices may provide fragment ions at low sensitivity, which make the identification process 575 difficult (Luz Gómez-Pérez et al., 2015). This aspect may have a greater impact on the 576 performances in the case of our presented study. Indeed, the FIT algorithm used for 577 identification takes into account all the expected fragment ions of a compound in the 578 calculation, and not only the most abundant fragment ion(s). At 0.01 mg.kg<sup>-1</sup>, some expected 579 fragment ions were missing in the experimental fragmentation spectra, which lowers the 580 581 similarity scores computed by the software. This could explained the difference in the performances observed in both studies. In addition, another factor should be highlighted. Luz 582 583 Gómez-Pérez et al. (2015) used DIA mode for their screening and identification. In our study, not all pesticides were fragmented with IDA-PIL mode, preventing them from being 584 585 validated. Thus, it would be necessary in a future work to evaluate DIA mode that allows all precursor ions to be fragmented. Nevertheless, the use of DIA mode would require 586 optimisation as it provides more interferences and complex spectra that reduce the confidence 587 588 into the identification process (Diallo et al., 2022).

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Regarding validation at the identification stage, the approach proposed in the present study is 590 not covered by the analytical guidelines, being limited by the options provided in 591 MasterView<sup>TM</sup> software. In this sense, manufacturers should be able to provide compliant 592 593 identification strategies regarding guidelines and the literature. In parallel, there is a need to improve existing official guidelines (SANTE, 2021; FDA, 2015) to include the identification 594 approach using MS/MS similarity in addition to the approach already proposed. Furthermore, 595 596 there are no requirements regarding false positives for method validation currently. There is also a need to indicate the false positive rate not to be exceeded to be able to use a screening 597 method in a high-throughput sample analysis-based context with a high confidence and to 598 avoid time-consuming manual verifications. 599

600 3.5. Suspect screening analysis

HRMS not only offers the possibility of conducting targeted analyses, but also of reprocessing
 retrospectively acquired data (Phillips et al., 2018), in order to broaden the analytical scope

and optimize monitoring in an exposomic-based context. Thus, the versatility and feasibility 603 of the method was also demonstrated performing post-acquisition suspect-screening analysis 604 on the analysed baby food samples. For this purpose, an expanded list of over 500 organic 605 contaminants including pesticides and metabolites, flame retardants and plasticizers, was 606 607 included in the database on the basis of their elemental formula and corresponding exact mass (Table S5). . Compound identification in suspect-screening analysis was conducted according 608 to the identification confidence levels proposed by Schymanski et al. (2014). In the first 609 instance, only the accurate mass measurement  $\pm 5$  ppm was used as a filter (confidence level 610 611 5). Among the positive results, if the compound was not fragmented, it was not considered anymore as a positive result as it was not feasible to identify the contaminant. If the 612 613 compound was fragmented, the resulting fragmentation spectrum was compared to the spectra available in online databases. If the experimental spectrum matched with external spectrum, 614 615 the compound was then considered as putatively identified (confidence level 2). Using this workflow, a flame retardant and plasticizer, tris(2-chloroethyl) phosphate (TCEP), was 616 617 putatively identified in some baby food samples. Afterwards, the analytical standard of TCEP was purchased and injected for confirmation. The criteria used were the  $t_r$  variation  $\pm 0.1$  min 618 619 and accurate mass measurement  $\pm$  5 ppm on the precursor ion, as well as the spectral similarity between the reference and experimental MS/MS fragmentation spectra, using FIT 620 algorithm with a similarity score > 70%. Thus, the presence of TCEP was confirmed at the 621 confidence level 1 of identification, as shown in Fig.4. The t<sub>r</sub> variation was 0.01 min, the mass 622 623 error was -2.5 ppm and the FIT score was 95.7%.

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The proposed method is a qualitative method acting as an alert system, i.e., it is able to detect and identify the presence or absence of a contaminant. Nevertheless, it would be important to quantify TCEP, either through a semi-quantitative approach, or precisely through conventional tandem mass spectrometry to determine whether its presence in the baby food samples exceeds the MRL.

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### 631 **4. Conclusion**

In this study, a QuEChERS-based protocol coupled to UHPLC-Q-TOF MS (ESI+) analysis enabled the screening of 204 pesticides in baby foods. The extraction procedure was optimized based on QuEChERS methods, to set-up a sample preparation protocol that can be adapted to various baby foods. The QuEChERS based on the AOAC method followed by a

purification using MgSO<sub>4</sub>/PSA, provided the less false negatives and the better signal-to-noise 636 ratio. Screening of pesticides in baby food samples was performed at the detection and 637 identification stages using a homemade tr-MS/MSlibrary, allowing for fast, automated and 638 unequivocal compound identification. As there is no standardisation, different thresholds on 639 the t<sub>r</sub> and accurate mass measurement were assessed on the performances of detection. 640 SANTE guidelines criteria ( $t_r \pm 0.1$  min, accurate mass measurement  $\pm 5$  ppm) showed the 641 best balance between false negatives with 16% and 5% at 0.01 mg.kg<sup>-1</sup> and 0.1 mg.kg<sup>-1</sup>, 642 respectively, and false positives with a low rate of 5%. Regarding identification, IDA and 643 IDA-PIL modes as well as different MS/MS algorithms-score combinations were evaluated. 644 IDA-PIL mode and identification using the FIT algorithm and similarity score >70% were 645 selected for the method, and demonstrated good performances with 37% and 15% false 646 negative rates, at 0.01 mg.kg<sup>-1</sup> and 0.1 mg.kg<sup>-1</sup>, respectively. The method was then validated 647 based on the SANTE guidelines for screening methods: 95% and 73% of the pesticides had a 648 screening SDL and LOI  $\leq 0.1$  mg.kg<sup>-1</sup>, respectively. The proposed method enabled the 649 650 unequivocal identification of the pesticide difenoconazole and a plasticizer, TCEP, in some of the investigated samples. 651

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Looking forward, the DIA mode should be further evaluated for pesticides and similar contaminants as it makes it possible to overcome the loss of MS/MS data and the related false negatives obtained in IDA mode. As the MS/MS spectral similarity is not yet integrated in analytical guidelines, there is a need to consider this type of identification.

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### 663 **Declaration of competing interest**

664 The authors declare that they have no conflicts of interest.

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## 666 **References**

- Aguilera-Luiz, M.M., Martínez Vidal, J.L., Romero-González, R., Garrido Frenich, A., 2012.
  Multiclass method for fast determination of veterinary drug residues in baby food by
  ultra-high-performance liquid chromatography-tandem mass spectrometry. Food
  Chemistry, 6th International Conference on Water in Food 132, 2171–2180.
  https://doi.org/10.1016/j.foodchem.2011.12.042
- Alder, L., Greulich, K., Kempe, G., Vieth, B., 2006. Residue analysis of 500 high priority pesticides: Better by GC–MS or LC–MS/MS? Mass Spectrometry Reviews 25, 838–
  865. https://doi.org/10.1002/mas.20091
- Anastassiades, M., Scherbaum, E., Taşdelen, B., Štajnbaher, D., 2007. Recent Developments
  in QuEChERS Methodology for Pesticide Multiresidue Analysis, in: Pesticide
  Chemistry. John Wiley & Sons, Ltd, pp. 439–458.
  https://doi.org/10.1002/9783527611249.ch46
- Anastassiades, M., Lehotay, S.J., Štajnbaher, D., Schenck, F.J., 2003. Fast and Easy
  Multiresidue Method Employing Acetonitrile Extraction/Partitioning and "Dispersive
  Solid-Phase Extraction" for the Determination of Pesticide Residues in Produce.
  Journal of AOAC INTERNATIONAL 86, 412–431.
  https://doi.org/10.1093/jaoac/86.2.412
- Diallo, T., Makni, Y., Lerebours, A., Thomas, H., Guérin, T., Parinet, J., 2022. Development
  and validation according to the SANTE guidelines of a QuEChERS-UHPLC-QTOFMS method for the screening of 204 pesticides in bivalves. Food Chemistry 386,
  132871. https://doi.org/10.1016/j.foodchem.2022.132871
- Dubocq, F., Kärrman, A., Gustavsson, J., Wang, T., 2021. Comprehensive chemical characterization of indoor dust by target, suspect screening and nontarget analysis using LC-HRMS and GC-HRMS. Environmental Pollution 276, 116701.
  https://doi.org/10.1016/j.envpol.2021.116701
- European Commission (2005). No 396/2005 of the European Parliament and of the Council of
   23 February 2005 on maximum residue levels of pesticides in or on food and feed of
   plant and animal origin and amending Council Directive 91/414/EEC
- European Commission. (2006). No 2006/125/EC of 5 December 2006 on processed cerealbased foods and baby foods for infants and young children
- European Commission. (2021). Analytical quality control and method validation procedurefor pesticide residues analysis in food and feed. SANTE/11312/2021.
- Food, Drug Administration, Office of Foods and Veterinary Medicine, Acceptance Criteria
   for Confirmation of Identity of Chemical Residues using Exact Mass Data.
- Gómez-Pérez, M.L., Romero-González, R., Vidal, J.L.M., Frenich, A.G., 2015. Identification
  of transformation products of pesticides and veterinary drugs in food and related
  matrices: Use of retrospective analysis. Journal of Chromatography A 1389, 133–138.
  https://doi.org/10.1016/j.chroma.2015.02.052
- Gómez-Ramos, M.M., Ferrer, C., Malato, O., Agüera, A., Fernández-Alba, A.R., 2013. 705 Liquid chromatography-high-resolution mass spectrometry for pesticide residue 706 analysis in fruit and vegetables: Screening and quantitative studies. Journal of 707 708 Chromatography Α, Editors' Choice VII 1287, 24-37. https://doi.org/10.1016/j.chroma.2013.02.065 709
- 710

González-Curbelo, M.Á., Socas-Rodríguez, B., Herrera-Herrera, A.V., González-Sálamo, J.,
 Hernández-Borges, J., Rodríguez-Delgado, M.Á., 2015. Evolution and applications of

- the QuEChERS method. TrAC Trends in Analytical Chemistry, Green Extraction
  Techniques 71, 169–185. https://doi.org/10.1016/j.trac.2015.04.012
- Guo, Z., Huang, S., Wang, J., Feng, Y.-L., 2020. Recent advances in non-targeted screening
  analysis using liquid chromatography high resolution mass spectrometry to explore
  new biomarkers for human exposure. Talanta 219, 121339.
  https://doi.org/10.1016/j.talanta.2020.121339
- Hercegová, A., Dömötörová, M., Matisová, E., 2007. Sample preparation methods in the analysis of pesticide residues in baby food with subsequent chromatographic determination. Journal of Chromatography A, Advances in Sample Preparation 1153, 54–73. https://doi.org/10.1016/j.chroma.2007.01.008
- Jia, W., Chu, X., Ling, Y., Huang, J., Chang, J., 2014. High-throughput screening of pesticide
   and veterinary drug residues in baby food by liquid chromatography coupled to
   quadrupole Orbitrap mass spectrometry. Journal of Chromatography A 1347, 122–
   128. https://doi.org/10.1016/j.chroma.2014.04.081
- Kim, L., Lee, D., Cho, H.-K., Choi, S.-D., 2019. Review of the QuEChERS method for the analysis of organic pollutants: Persistent organic pollutants, polycyclic aromatic hydrocarbons, and pharmaceuticals. Trends in Environmental Analytical Chemistry 22, e00063. https://doi.org/10.1016/j.teac.2019.e00063
- 731 Lehotay, S.J., Collaborators:, 2007. Determination of Pesticide Residues in Foods by
  732 Acetonitrile Extraction and Partitioning with Magnesium Sulfate: Collaborative Study.
  733 Journal of AOAC INTERNATIONAL 90, 485–520.
  734 https://doi.org/10.1093/jaoac/90.2.485
- Li, H., Wu, J., Chen, C., Xin, W., Zhang, W., 2021. Simultaneous determination of 74 pesticide residues in Panax notoginseng by QuEChERS coupled with gas chromatography tandem mass spectrometry. Food Science and Human Wellness 10, 241–250. https://doi.org/10.1016/j.fshw.2021.02.014
- López, M.G., Fussell, R.J., Stead, S.L., Roberts, D., McCullagh, M., Rao, R., 2014.
  Evaluation and validation of an accurate mass screening method for the analysis of
  pesticides in fruits and vegetables using liquid chromatography–quadrupole-time of
  flight–mass spectrometry with automated detection. Journal of Chromatography A
  1373, 40–50. https://doi.org/10.1016/j.chroma.2014.10.099
- Lukowicz, C., Ellero, -Simatos Sandrine, R, égnier M., Polizzi, A., Lasserre, F., Montagner,
  A., Lippi, Y., Jamin, E.L., Martin, J.-F., Naylies, C., Canlet, C., Debrauwer, L.,
  Bertrand, -Michel Justine, Al, S.T., Th, éodorou V., Loiseau, N., Mselli, -Lakhal
  Laïla, Guillou, H., Gamet, -Payrastre Laurence, n.d. Metabolic Effects of a Chronic
  Dietary Exposure to a Low-Dose Pesticide Cocktail in Mice: Sexual Dimorphism and
  Role of the Constitutive Androstane Receptor. Environmental Health Perspectives
  126, 067007. https://doi.org/10.1289/EHP2877
- Luz Gómez-Pérez, M., Romero-González, R., José Luis Martínez, V., Garrido Frenich, A.,
   2015. Analysis of pesticide and veterinary drug residues in baby food by liquid
   chromatography coupled to Orbitrap high resolution mass spectrometry. Talanta 131,
   1–7. https://doi.org/10.1016/j.talanta.2014.07.066
- Musarurwa, H., Chimuka, L., Pakade, V.E., Tavengwa, N.T., 2019. Recent developments and 755 756 applications of QuEChERS based techniques on food samples during pesticide Journal Food Composition and 757 analysis. of Analysis 84. 103314. https://doi.org/10.1016/j.jfca.2019.103314 758
- Nougadère, A., Sirot, V., Cravedi, J.-P., Vasseur, P., Feidt, C., Fussell, R.J., Hu, R., Leblanc,
  J.-C., Jean, J., Rivière, G., Sarda, X., Merlo, M., Hulin, M., 2020. Dietary exposure to
  pesticide residues and associated health risks in infants and young children Results

- of the French infant total diet study. Environment International 137, 105529.
  https://doi.org/10.1016/j.envint.2020.105529
- 764 Ntzani, E.E., Ntritsos G, C.M., Evangelou, E., Tzoulaki, I., 2013. Literature review on
  765 epidemiological studies linking exposure to pesticides and health effects. EFSA
  766 Supporting Publications 10, 497E. https://doi.org/10.2903/sp.efsa.2013.EN-497
- 767 Ogawa, T., Zaitsu, K., Kokaji, T., Suga, K., Kondo, F., Iwai, M., Suzuki, T., Ishii, A., Seno, H., 2020. Development and application of a forensic toxicological library for 768 identification of 56 natural toxic substances by liquid chromatography-quadrupole 769 time-of-flight mass spectrometry. Forensic Toxicol 232 - 242.770 38. 771 https://doi.org/10.1007/s11419-019-00506-w
- Pang, X., Liu, X., Peng, L., Chen, Z., Qiu, J., Su, X., Yu, C., Zhang, J., Weng, R., 2021.
  Wide-scope multi-residue analysis of pesticides in beef by ultra-high-performance
  liquid chromatography coupled with quadrupole time-of-flight mass spectrometry.
  Food Chemistry 351, 129345. https://doi.org/10.1016/j.foodchem.2021.129345
- Pérez-Ortega, P., Lara-Ortega, F.J., Gilbert-López, B., Moreno-González, D., García-Reyes,
  J.F., Molina-Díaz, A., 2017. Screening of Over 600 Pesticides, Veterinary Drugs,
  Food-Packaging Contaminants, Mycotoxins, and Other Chemicals in Food by UltraHigh Performance Liquid Chromatography Quadrupole Time-of-Flight Mass
  Spectrometry (UHPLC-QTOFMS). Food Anal. Methods 10, 1216–1244.
  https://doi.org/10.1007/s12161-016-0678-0
- Petrarca, M.H., Braga, P.A. de C., Reyes, F.G.R., Bragotto, A.P.A., 2022. Exploring
  miniaturized sample preparation approaches combined with LC-QToF-MS for the
  analysis of sulfonamide antibiotic residues in meat- and/or egg-based baby foods.
  Food Chemistry 366, 130587. https://doi.org/10.1016/j.foodchem.2021.130587
- Phillips, K.A., Yau, A., Favela, K.A., Isaacs, K.K., McEachran, A., Grulke, C., Richard,
  A.M., Williams, A.J., Sobus, J.R., Thomas, R.S., Wambaugh, J.F., 2018. Suspect
  Screening Analysis of Chemicals in Consumer Products. Environ. Sci. Technol. 52,
  3125–3135. https://doi.org/10.1021/acs.est.7b04781
- Rizzati, V., Briand, O., Guillou, H., Gamet-Payrastre, L., 2016. Effects of pesticide mixtures
  in human and animal models: An update of the recent literature. Chemico-Biological
  Interactions 254, 231–246. https://doi.org/10.1016/j.cbi.2016.06.003
- Schymanski, E.L., Jeon, J., Gulde, R., Fenner, K., Ruff, M., Singer, H.P., Hollender, J.,
  2014a. Identifying Small Molecules via High Resolution Mass Spectrometry:
  Communicating Confidence. Environ. Sci. Technol. 48, 2097–2098.
  https://doi.org/10.1021/es5002105
- 797 Schymanski, E.L., Jeon, J., Gulde, R., Fenner, K., Ruff, M., Singer, H.P., Hollender, J.,
  798 2014b. Identifying Small Molecules via High Resolution Mass Spectrometry:
  799 Communicating Confidence. Environ. Sci. Technol. 48, 2097–2098.
  800 https://doi.org/10.1021/es5002105
- Villaverde, J.J., Sevilla-Morán, B., López-Goti, C., Alonso-Prados, J.L., Sandín-España, P.,
  2016. Trends in analysis of pesticide residues to fulfil the European Regulation (EC)
  No. 1107/2009. TrAC Trends in Analytical Chemistry 80, 568–580.
  https://doi.org/10.1016/j.trac.2016.04.017
- Walorczyk, S., Drożdżyński, D., Kierzek, R., 2015. Determination of pesticide residues in samples of green minor crops by gas chromatography and ultra performance liquid chromatography coupled to tandem quadrupole mass spectrometry. Talanta 132, 197–204. https://doi.org/10.1016/j.talanta.2014.08.073
- Wang, J., Chow, W., Wong, J.W., Leung, D., Chang, J., Li, M., 2019. Non-target data
  acquisition for target analysis (nDATA) of 845 pesticide residues in fruits and

- vegetables using UHPLC/ESI Q-Orbitrap. Anal Bioanal Chem 411, 1421–1431.
  https://doi.org/10.1007/s00216-019-01581-z
- Wang, Z., Cao, Y., Ge, N., Liu, X., Chang, Q., Fan, C., Pang, G.-F., 2016. Wide-scope screening of pesticides in fruits and vegetables using information-dependent acquisition employing UHPLC-QTOF-MS and automated MS/MS library searching.
  Anal Bioanal Chem 408, 7795–7810. https://doi.org/10.1007/s00216-016-9883-3
- Weng, R., Lou, S., Pang, X., Song, Y., Su, X., Xiao, Z., Qiu, J., 2020. Multi-residue analysis
  of 126 pesticides in chicken muscle by ultra-high-performance liquid chromatography
  coupled to quadrupole time-of-flight mass spectrometry. Food Chemistry 309,
  125503. https://doi.org/10.1016/j.foodchem.2019.125503
- Wild, C.P., 2005. Complementing the Genome with an "Exposome": The Outstanding
  Challenge of Environmental Exposure Measurement in Molecular Epidemiology.
  Cancer Epidemiol Biomarkers Prev 14, 1847–1850. https://doi.org/10.1158/10559965.EPI-05-0456
- Zhang, Z., Dong, M., Hao, X., Han, L., Song, S., Yao, W., 2019. Evaluation of cleanup
  procedures in pesticide multi-residue analysis with QuEChERS in cinnamon bark.
  Food Chemistry 276, 140–146. https://doi.org/10.1016/j.foodchem.2018.10.019
- 828 829

# 830 Figure captions

Fig.1. Comparison of different QuEChERS-based procedures for sample preparationoptimisation.

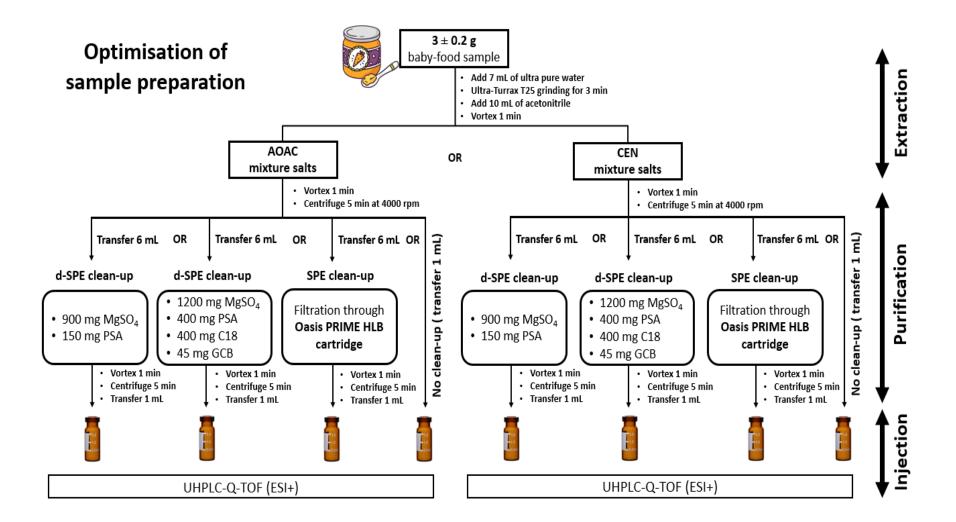
833 Fig.2. Impact of different extraction and purification procedures on the false negative rates

(%) obtained for (a) baby food 1; (b) baby food 2; (c) baby food 3; and (d) baby food 4. The

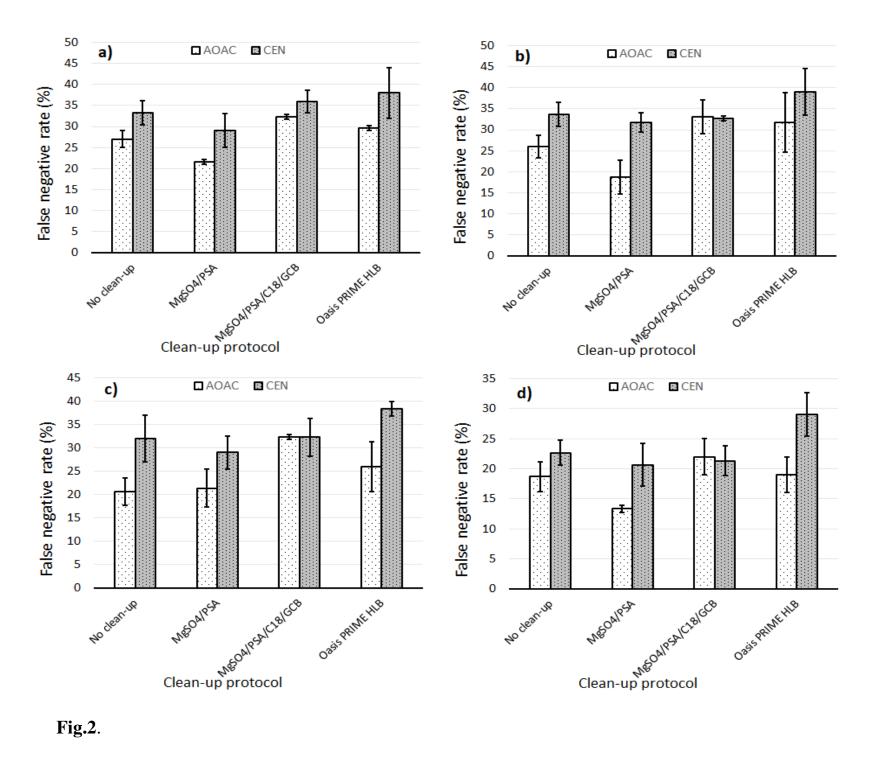
error bars correspond to the standard deviation for the response measured (Mean  $\pm$  SD, n=3).

**Fig.3**. Unexpected identification of difenoconazole in non-spiked baby food 5, with corresponding (a) extracted ion chromatogram; (b) isotopic pattern; and (c) experimental and reference fragmentation spectra comparison using FIT algorithm.

- **Fig.4**. Identification of tris(2-chloroethyl) phosphate in non-spiked baby food 1, with corresponding (a) extracted ion chromatogram; (b) isotopic pattern; (c) experimental and reference fragmentation spectra comparison using FIT algorithm.
- 842



**Fig.1**.





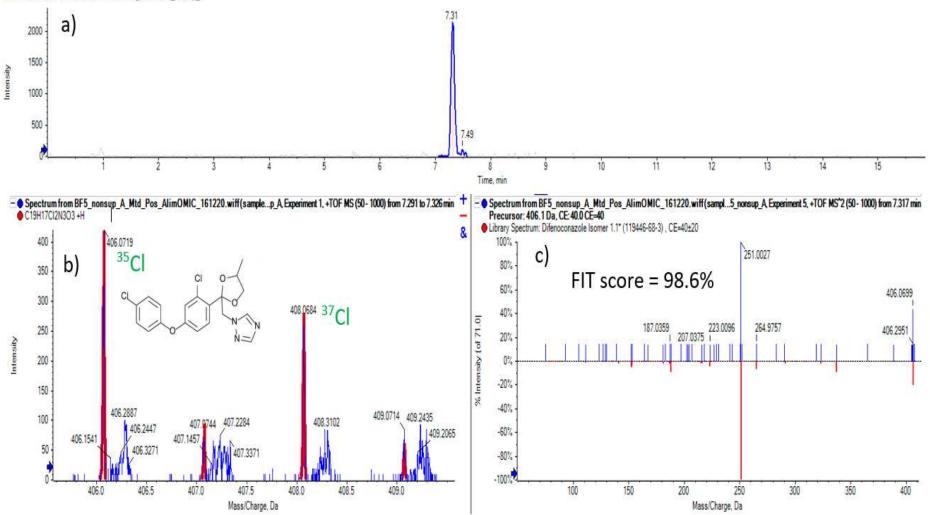
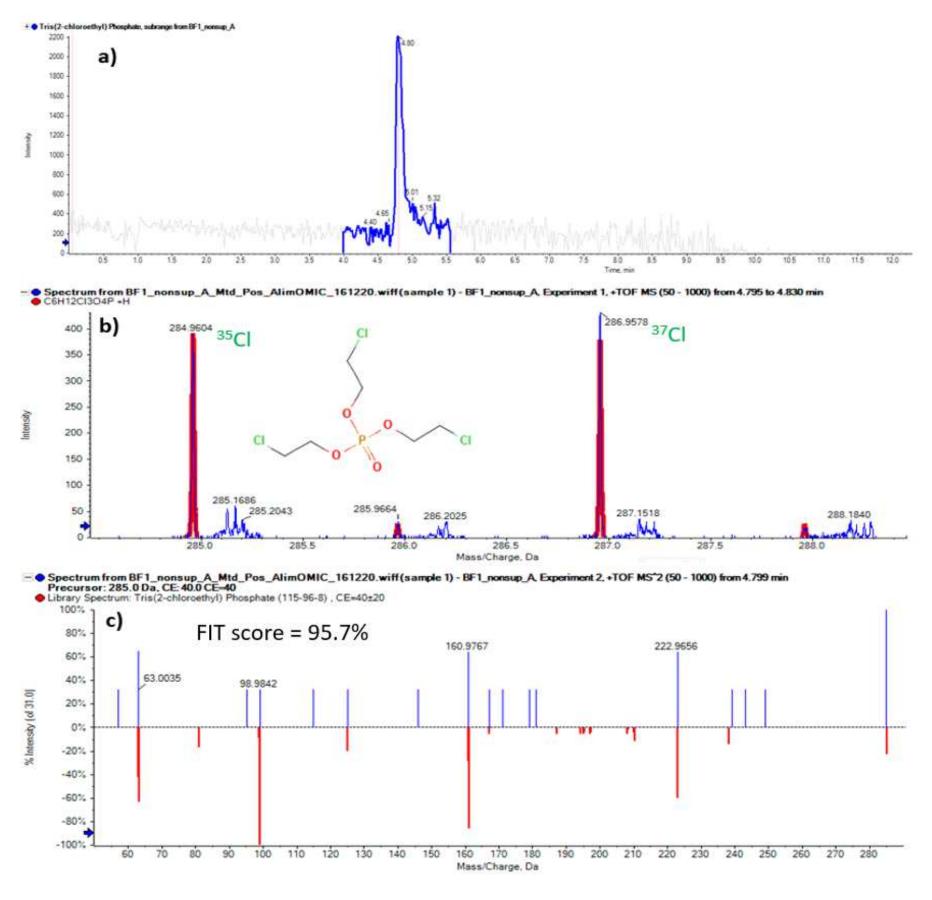


Fig.3.





**Table 1**: Impact of different thresholds on the  $t_r$  and accurate mass measurement criteria regarding the false negatives (%), in a set of 10 baby food samples in triplicate, at 0.01 and 0.1 mg.kg<sup>1</sup>.

	False negative (%)* ± SD (%)         Reported criteria for detection						
	$(t_r \pm 0.1 \text{ min}; \text{ME} \pm 5\text{-ppm})^{[1]}$		$(t_r \pm 0.5 \text{ min}; \text{ME} \pm 5\text{-ppm})^{[2]}$		$(t_r \pm 0.25 \text{ min; ME} \pm 10\text{-ppm})^{[3]}$		
-	Concentration (mg.kg <sup>-1</sup> )						
Sample (n=3)	0.01	0.1	0.01	0.1	0.01	0.1	
Baby food 5	<b>16</b> ± 1	<b>6</b> ± 2	<b>16</b> ± 1	<b>6</b> ± 1	<b>12 ±</b> 2	<b>4</b> ± 1	
Baby food 6	<b>17</b> ± 2	<b>5</b> ± 2	<b>17 ±</b> 2	<b>5</b> ± 2	<b>11 ±</b> 2	<b>4</b> ± 2	
Baby food 7	<b>15</b> ± 4	<b>9</b> ± 7	<b>14</b> ± 4	<b>9</b> ±7	<b>12</b> ± 3	<b>3</b> ±1	
Baby food 8	<b>21</b> ± 1	<b>5</b> ± 1	<b>20</b> ± 1	<b>5</b> ± 1	<b>12</b> ± 3	<b>3</b> ±1	
Baby food 9	<b>17</b> ± 5	<b>6</b> ± 2	<b>16 ±</b> 5	<b>6</b> ± 2	<b>10</b> ± 3	<b>4</b> ± 2	
Baby food 10	<b>16</b> ± 7	<b>5</b> ± 1	<b>15</b> ± 7	<b>5</b> ± 1	11 ± 5	<b>3</b> ±1	
Baby food 11	<b>15</b> ± 5	<b>5</b> ± 2	<b>14 ±</b> 4	<b>5</b> ± 2	<b>9</b> ±5	<b>3</b> ±1	
Baby food 12	<b>13</b> ± 4	<b>4</b> ± 0	<b>13 ±</b> 1	$4 \pm 0$	<b>9</b> ± 3	$3 \pm 0$	
Baby food 13	<b>17</b> ± 6	<b>5</b> ± 2	<b>16</b> ± 6	<b>5</b> ± 2	<b>13</b> ± 4	<b>3</b> ± 1	
Baby food 14	<b>15</b> ± 4	<b>3</b> ±0	<b>15</b> ± 2	<b>3</b> ±0	<b>10</b> ± 4	$3 \pm 0$	
Average (n = 30) ± SD (%)	<b>16 ±</b> 4	<b>5</b> ± 3	<b>16 ±</b> 4	<b>5</b> ± 3	<b>11 ±</b> 3	<b>3</b> ±1	

<sup>[1]</sup> SANTE/12682/2019; <sup>[2]</sup> FDA, 2015; Wang et al., 2019; <sup>[3]</sup> Pérez-Ortega et al., 2017

\* The percentages were calculated based on 199 compounds as some pesticides had identical formulas and were co-eluted: aldicarb sulfone/butoxycarboxim ( $C_7H_{14}N_2O_4S$ ); desmedipham/phenmedipham ( $C_{16}H_{16}N_2O_4$ ); prometon/secbumeton/terbumeton ( $C_{10}H_{19}N_5O$ ) and prometryne/terbutryn ( $C_{10}H_{19}N_5S$ )

		False negative $(\%)^* \pm SD(\%)$				
		Concentration (mg.kg <sup>-1</sup> )				
		<b>0.01 (MRL)</b>		0.1		
MS/MS Algorithm	Similarity score	Method A	Method B	Method A	Method B	
PUR	>50%	<b>52</b> ± 8	<b>36</b> ± 7	<b>27</b> ± 4	<b>14 ±</b> 1	
PUR	>70%	<b>55</b> ± 8	<b>42</b> ± 8	<b>28</b> ± 4	<b>17</b> ± 2	
FIT	>50%	<b>50</b> ± 8	<b>33</b> ± 7	<b>26</b> ± 4	<b>14 ±</b> 1	
FIT	>70%	<b>51</b> ± 8	<b>37</b> ± 8	<b>27</b> ± 4	<b>15 ±</b> 2	
RFIT	>50%	<b>50</b> ± 9	<b>29</b> ± 6	<b>26</b> ± 4	<b>13</b> ± 1	
RFIT	>70%	<b>52</b> ± 8	<b>32</b> ±6	<b>26</b> ± 4	<b>14 ±</b> 1	

**Table 2**: Impact of a PIL and different algorithm-score combinations on the false negatives (%), in a set of 10 baby food samples in triplicate, at 0.01 and 0.1 mg.kg<sup>-1</sup>.

\* The percentages were calculated based on 199 compounds as some pesticides had identical formulas and were co-eluted: aldicarb sulfone/butoxycarboxim ( $C_7H_{14}N_2O_4S$ ); desmedipham/phenmedipham ( $C_{16}H_{16}N_2O_4$ ); prometon/secbumeton/terbumeton ( $C_{10}H_{19}N_5O$ ) and prometryne/terbutryn ( $C_{10}H_{19}N_5S$ )